Science Advances NAAAS

advances.sciencemag.org/cgi/content/full/2/10/e1600842/DC1

Supplementary Materials for

Climatic warming destabilizes forest ant communities

Sarah E. Diamond, Lauren M. Nichols, Shannon L. Pelini, Clint A. Penick, Grace W. Barber, Sara Helms Cahan, Robert R. Dunn, Aaron M. Ellison, Nathan J. Sanders, Nicholas J. Gotelli

> Published 28 October 2016, *Sci. Adv.* **2**, e1600842 (2016) DOI: 10.1126/sciadv.1600842

This PDF file includes:

- Levins' metapopulation, turnover, and persistence models
- Alternative demographic and transition matrix model specifications
- Linking altered community dynamics with changes in community composition
- fig. S1. Frequency of nest box censuses at the two experimental warming arrays.
- fig. S2. Mean proportion of nest boxes occupied per chamber at Duke Forest.
- fig. S3. Mean proportion of nest boxes colonized per chamber at Duke Forest.
- fig. S4. Mean proportion of nest boxes that went extinct per chamber at Duke Forest.
- fig. S5. Mean proportion of nest boxes occupied per chamber at Harvard Forest.
- fig. S6. Mean proportion of nest boxes colonized per chamber at Harvard Forest.
- fig. S7. Mean proportion of nest boxes that went extinct per chamber at Harvard Forest.
- fig. S8. Equilibrium frequencies as a function of chamber temperature for each of the four focal species and empty nest boxes at Duke Forest.
- fig. S9. Equilibrium frequencies as a function of chamber temperature for each of the five focal species and empty nest boxes at Harvard Forest.
- fig. S10. Mean proportion of nest boxes occupied at equilibrium using Levins' colonization-extinction formula at Duke Forest.
- fig. S11. Mean proportion of nest boxes that turn over at Duke Forest.
- fig. S12. Mean proportion of nest boxes occupied at equilibrium using Levins' colonization-extinction formula at Harvard Forest.
- fig. S13. Mean proportion of nest boxes that turn over at Harvard Forest.
- fig. S14. Mean proportion of nest boxes that persisted to the next census per chamber at Duke Forest.
- fig. S15. Mean proportion of nest boxes that persisted to the next census per chamber at Harvard Forest.
- table S1. Observed transitions in the nest boxes at Duke and Harvard forests.
- table S2. Models of species associations at Duke and Harvard forests.
- table S3. Models of temperature effects at Duke Forest.
- table S4. Models of temperature effects at Harvard Forest.
- table S5. Models of community-wide responses.
- table S6. Models of equilibrium frequency as functions of temperature.
- table S7. Temperature dependence of individual transition probabilities.
- table S8. Transition matrix correlates of community stability.

Supplementary Materials

Levins' metapopulation, turnover, and persistence models

At Duke Forest, we found a strong concordance between the relationship of raw occupancy as a function of chamber temperature and equilibrium occupancy based on the Levins' metapopulation model. Like the raw occupancy results, the equilibrium occupancy showed *Crematogaster lineolata* significantly increasing with increasing chamber temperature, and a trend toward *Aphaenogaster* spp. decreasing with increasing temperature (fig. S12 and table S3). Turnover was largely insensitive to temperature except for *Brachyponera chinensis*, which exhibited significantly reduced turnover with increasing temperature; however, *B. chinensis* had overall few occurrences in the chambers, and more data are needed to confirm this result (fig. S13 and table S3).

At Harvard Forest, we found little concordance between the relationship of raw occupancy as a function of chamber temperature and equilibrium occupancy based on the Levins' metapopulation model (fig. S14 and table S4). None of the relationships for any of the five focal taxa was significant, however, we note that there were fewer census points available for Harvard Forest compared with Duke Forest, limiting our ability to estimate these effects. Further, it is possible the Harvard Forest ant community was not at equilibrium, which violates the assumptions of the Levins' model, and could account for this difference between the results of the raw occupancy and the equilibrium occupancy as functions of chamber temperature. Turnover was insensitive to temperature for each of the five focal taxa at Harvard Forest (fig. S₁₅ and table S₄).

Although we observed weak trends of increasing persistence among thermophilic species in warmer environments (*i.e.*, *C. lineolata*, *Temnothorax longispinosus*, and *Camponotus* spp.), persistence in nest boxes was not significantly affected by chamber temperature for any of the species at either Duke Forest (fig. S14 and table S3) or Harvard Forest (fig. S15 and table S4); however, this result may reflect the high replication needed to detect persistence (see Materials and Methods).

Alternative demographic and transition matrix model specifications

Our primary analyses considered nest boxes to be empty for the focal species when occupied by a non-focal species. Alternatively, we can consider nest boxes to be unavailable for the focal species when occupied by a non-focal species. We refit our models of colonization, extinction, occupancy and persistence treating nest boxes as not available when a non-focal species inhabited the nest box, and found the results to be qualitatively similar to our focal analyses which treat nest boxes inhabited by the non-focal species as empty for the focal species.

To isolate the effects of warming on species interactions that involved usurpation of nest boxes (*i.e.*, as a subset of all species interactions ranging from usurpation of nest boxes to competition for resources near the nest sites), we recalculated the proportion of colonization and extinction events as those involving a change from a non-focal species to a focal species (colonization event) and vice versa (extinction event) in back-to-back census points. Although the number of colonization and extinction events were fewer using this definition, our findings were qualitatively similar to our main analyses using a broader definition of colonization and extinction that allowed for intervening periods when nest boxes were empty. Together, these results suggest that the mechanism of altered species interactions via nest usurpation plays an important role in the shifts in colonization and extinction we observed under experimental warming.

Although our census intervals were relatively frequent, on the order of monthly during the growing season, it is possible that transient colonization and extinction occurred between sampling intervals. We refit our Markov transition models where we replaced 0 count cells with a value of 1 to explore the effects of imperfect detection. This change created a small transition probability where originally no transition was permitted with a value of 0. We repeated our analyses of damping ratio, equilibrium frequencies for each species and the empty state for nest boxes, and the individual transition probabilities, and found our results to be qualitatively similar to our original analyses.

We also refit our Markov transition models with different numbers of focal species: initially we had 5 focal species (or species groups) at Harvard Forest, and refit our Markov models using the 4 and 3 most abundant species. We employed the same procedure for Duke Forest: initially we had 4 focal species (or species groups), and refit our Markov models using the 3 most abundant species. Regardless of the number of species included at either site, we found qualitatively similar results using the reduced numbers of focal species. Our results for occupancy, colonization, extinction and community stability therefore appear to be robust to potential issues with sampling and detection.

Linking altered community dynamics with changes in community composition

Our analyses explored the dynamical stability of communities under climatic change, and found evidence to suggest a loss of stability with increasing environmental temperature. Much more frequently when climate researchers quantify the effects of climatic warming on communities, metrics such as species abundances and community composition are reported, often for a single or limited number of time points (*26*). It is reasonable then to ask how well the results for dynamical stability align with these more commonly used metrics of community responses to climatic change (and relatedly, whether the dynamics of interactions among individuals in and around nest boxes scale up to higher-level community responses). At least in the forest ant communities we monitored, there appears to be a fairly strong congruence between the loss of stability with increasing environmental temperature, and shifts in individual species abundances and community composition. Across each of these metrics, thermophilic species appear to be driving ant community responses to warming: thermophilic species occupy nest sites for extended periods of time which, as our work in this study suggests, destabilizes communities; the abundances of thermophilic species tend to increase (*13*); and thermophilic species appear to drive compositional shifts in communities (*20*). Interestingly, there is more of a discrepancy among these metrics when we take geographic location into account. Despite the strong impact of warming on dynamical stability at both our lower (warmer) and higher (cooler) latitude sites, the impacts of warming on species abundance and community composition were much stronger at our lower latitude (warmer) site. Understanding the connections between dynamical community stability and community compositional changes under warming across different geographic locations is a prime area for future research efforts on biodiversity forecasting.

Supplementary Figures and Tables

fig. S1. Frequency of nest box censuses at the two experimental warming arrays. Nest box census dates are given as the Julian dates from 2011 through 2015 for Duke Forest (orange) and Harvard Forest (green).

fig. S2. Mean proportion of nest boxes occupied per chamber at Duke Forest. Dashed lines indicate the fitted quasi-binomial generalized linear model: see table S3 for model coefficients and statistics.

fig. S3. Mean proportion of nest boxes colonized per chamber at Duke Forest. Dashed lines indicate the fitted quasi-binomial generalized linear model: see table S3 for model coefficients and statistics.

fig. S4. Mean proportion of nest boxes that went extinct per chamber at Duke Forest. Dashed lines indicate the fitted quasi-binomial generalized linear model: see table S3 for model coefficients and statistics.

fig. S5. Mean proportion of nest boxes occupied per chamber at Harvard Forest. Dashed lines indicate the fitted quasi-binomial generalized linear model: see table S4 for model coefficients and statistics.

fig. S6. Mean proportion of nest boxes colonized per chamber at Harvard Forest. Dashed lines indicate the fitted quasi-binomial generalized linear model: see table S4 for model coefficients and statistics.

fig. S7. Mean proportion of nest boxes that went extinct per chamber at Harvard Forest.

Dashed lines indicate the fitted quasi-binomial generalized linear model: see table S4 for model coefficients and statistics.

fig. S8. Equilibrium frequencies as a function of chamber temperature for each of the four focal species and empty nest boxes at Duke Forest. Dashed lines indicate the fitted quasibinomial generalized linear model: see table S6 for model coefficients and statistics.

fig. S9. Equilibrium frequencies as a function of chamber temperature for each of the five focal species and empty nest boxes at Harvard Forest. Dashed lines indicate the fitted quasibinomial generalized linear model: see table S6 for model coefficients and statistics.

fig. S10. Mean proportion of nest boxes occupied at equilibrium using Levins' colonizationextinction formula at Duke Forest. Dashed lines indicate the fitted quasi-binomial generalized linear model: see table S3 for model coefficients and statistics.

fig. S11. Mean proportion of nest boxes that turn over at Duke Forest. Dashed lines indicate the fitted quasi-binomial generalized linear model: see table S3 for model coefficients and statistics.

fig. S12. Mean proportion of nest boxes occupied at equilibrium using Levins' colonizationextinction formula at Harvard Forest. Dashed lines indicate the fitted quasi-binomial generalized linear model: see table S4 for model coefficients and statistics.

fig. S13. Mean proportion of nest boxes that turn over at Harvard Forest. Dashed lines indicate the fitted quasi-binomial generalized linear model: see table S4 for model coefficients and statistics.

fig. S14. Mean proportion of nest boxes that persisted to the next census per chamber at Duke Forest. Dashed lines indicate the fitted quasi-binomial generalized linear model: see table S3 for model coefficients and statistics.

fig. S15. Mean proportion of nest boxes that persisted to the next census per chamber at Harvard Forest. Dashed lines indicate the fitted quasi-binomial generalized linear model: see table S4 for model coefficients and statistics.

table S1. Observed transitions in the nest boxes at Duke and Harvard forests. The number of transitions for a given state out of the total number of transitions for the transition matrices of each of the 15 chambers at each of the two sites.

table S2. Models of species associations at Duke and Harvard forests. Slope estimates ± 1 S.E. from quasi-binomial generalized linear models for the chamber temperature and species association models. *F*-ratios and *P*-values indicate the statistical significance of chamber temperature (MAT) or species associations (abbreviated with the first letter of the genus and first letter of the species name).

table S3. Models of temperature effects at Duke Forest. Slope estimates \pm 1 S.E. from quasibinomial generalized linear models for occupancy, colonization, extinction, persistence, equilibrium occupancy, and turnover as functions of chamber temperature at Duke Forest. *F*-ratios and *P*-values indicate the statistical significance of chamber temperature.

table S4. Models of temperature effects at Harvard Forest. Slope estimates ± 1 S.E. from quasi-binomial generalized linear models for occupancy, colonization, extinction, persistence, equilibrium occupancy, and turnover as functions of chamber temperature at Harvard Forest. *F*-ratios and *P*-values indicate the statistical significance of chamber temperature.

table S5. Models of community-wide responses. Slope estimates \pm 1 S.E. from generalized linear models using a quasi-binomial error structure to examine the main and interactive effects of mean annual temperature and site on community-wide occupancy, colonization, and extinction. *F*-ratios and *P*-values indicate the statistical significance of chamber temperature, site, and their interaction. The pseudo r^2 is calculated as $1 -$ (residual deviance/null deviance).

table S6. Models of equilibrium frequency as functions of temperature. Slope estimates ± 1 S.E. from quasi-binomial generalized linear models for equilibrium frequency as a function of chamber temperature at Duke Forest and Harvard Forest. *F*-ratios and *P*-values indicate the statistical significance of chamber temperature.

table S7. Temperature dependence of individual transition probabilities. Slope estimates ± 1 S.E. from quasi-binomial generalized linear models for individual species and empty nest box state transition probabilities as a function of chamber temperature at Duke Forest and Harvard Forest. *F*-ratios and *P*-values indicate the statistical significance of chamber temperature.

Site	Transition	Estimate	SE	\boldsymbol{F}	\boldsymbol{P}
Duke Forest	Empty to Empty	-0.0395	0.0385	1.05	0.324
	Empty to Aphaenogaster	-0.0467	0.0525	0.797	0.388
	Empty to Brachyponera	NA	NA	NA	NA
	Empty to Crematogaster	0.188	0.0903	4.46	0.0545
	Empty to Temnothorax	-0.0192	0.113	0.0289	0.87
	Aphaenogaster to Empty	0.0435	0.0737	0.348	0.565
	Aphaenogaster to Aphaenogaster	-0.0845	0.0739	1.32	0.272
	Aphaenogaster to Brachyponera	NA	NA	NA	NA
	Aphaenogaster to Crematogaster	0.261	0.173	2.43	0.143
	Aphaenogaster to Temnothorax	4.64E-15	0.228	$\boldsymbol{0}$	$\mathbf{1}$
	<i>Brachyponera</i> to Empty	NA	NA	NA	NA
	Brachyponera to Aphaenogaster	NA	NA	NA	NA
	Brachyponera to Brachyponera	NA	NA	NA	NA
	Brachyponera to Crematogaster	NA	NA	NA	NA
	Brachyponera to Temnothorax	NA	NA	NA	NA
	Crematogaster to Empty	-0.0936	0.152	0.38	0.548
	Crematogaster to Aphaenogaster	-0.444	0.179	7.78	0.0153
	Crematogaster to Brachyponera	NA	NA	NA	NA
	Crematogaster to Crematogaster	0.216	0.153	2.06	0.175
	Crematogaster to Temnothorax	-1.47	0.552	19.5	0.00309
	Temnothorax to Empty	-0.0886	0.41	0.0467	0.835
	Temnothorax to Aphaenogaster	-1.35	0.933	4.55	0.0705
	Temnothorax to Brachyponera	NA	NA	NA	NA
	Temnothorax to Crematogaster	-0.205	0.505	0.167	0.695
	Temnothorax to Temnothorax	0.573	0.376	2.76	0.14
Harvard Forest	Empty to Empty	-0.00984	0.0526	0.00042	0.855
	Empty to Aphaenogaster	-0.0184	0.0699	0.000748	0.796
	Empty to Camponotus	0.225	0.227	0.0118	0.347
	Empty to Lasius	-0.0339	0.277	0.000201	0.922
	Empty to Myrmica	-0.0279	0.103	0.00127	0.79
	Empty to Temnothorax	-0.0739	0.156	0.000603	0.67
	Aphaenogaster to Empty	0.0622	0.157	0.0397	0.697
	Aphaenogaster to Aphaenogaster	0.159	0.145	0.211	0.294
	Aphaenogaster to Camponotus	-0.0324	0.235	0.00145	0.896
	Aphaenogaster to Lasius	2.72E-15	0.49	$\boldsymbol{0}$	1
	Aphaenogaster to Myrmica	-1	0.755	1.29	0.0741
	Aphaenogaster to Temnothorax	$-7.49E-15$	0.506	$\boldsymbol{0}$	$\mathbf{1}$

table S8. Transition matrix correlates of community stability. Spearman's rank correlations between damping ratio and transition matrix attributes for Duke Forest and Harvard Forest.